

## **Branched and linear forms of PFAS – A means of a more comprehensive assessment of environmental impacts**

*Patrick van Hees, Eurofins Environment Testing Sweden AB*

### **Abbreviations:**

ECF = Electrochemical fluorination; N-EtFOSA = N-ethylperfluorooctanesulfonamide; N-EtFOSE = N-ethylperfluorooctanesulfonamidoethanol; N-MeFOSA = N-methylperfluorooctanesulfonamide; N-MeFOSE = N-methylperfluorooctanesulfonamidoethanol PFAS = Poly- and perfluoroalkyl substances; PFHxS = Perfluorohexane sulfonate; PFOA = Perfluorooctanoate; PFOS = Perfluorooctane sulfonate; PFOSA = Perfluorooctanesulfonamide

### *Introduction*

Over the last 20-25 years the research on both environmental and health related effects of PFAS has been intense. During the last 5-10 years the difference between isomeric forms of the substances has emerged as an additional factor. The physicochemical properties of the linear and branched forms vary slightly which leads to significant differences in biochemical reactions/transformations, bioaccumulation and thus potential toxic exposure. Adsorption onto solid phases is another property affected. All these factors also influence the relative distribution in environmental compartments such as water and sediment. The research has also created an opportunity to develop environmental forensic applications.

### *Manufacturing processes*

Electrochemical fluorination (ECF) and telomerisation are the two major methods employed to produce PFAS. The branched isomers of PFAS are mainly manufactured in the ECF method, which has historically been used to produce the major part of the two dominant PFAS, PFOS and PFOA. ECF gives rise to complex mixtures of linear and branched compounds. PFOA produced by this method has typically had an isomer composition of 78% linear (n-PFOA) and 22% branched isomers (br-PFOA). ECF-PFOS shows a distribution of around 70% linear (n-PFOS) and 30% branched (br-PFOS). Precursors and similar substances e.g. N-methyl and N-ethyl perfluorooctane sulfonamido ethanols (Me- and EtFOSEs) and sulfonamides (Me- and EtFOSAs) have ratios in the same range. On the other hand the telomerisation process keeps the structure of the starting telogen and a pure linear or isopropyl form is produced (Benskin et al., 2010; Jiang et al, 2015).

### *Distribution in environmental matrices*

The differing properties of linear and branched PFAS isomers can affect the partitioning and migration of the substances in the aquatic environment. A number of studies have reported enrichment of br-PFOS in natural waters as compared to technical products (Benskin et al., 2010; Chen et al., 2015 and references within).

In a nationwide study in France it was found that n-PFOS was the dominant isomer in sediments (88% of total; n=129), while in water n-PFOS comprised 48% (n=333) (Munoz et al., 2015). A similar distribution was observed by Houde et al. (2008). Chen et al. (2015) studied PFOS, PFOA and PFOSA in water, water particulate phase and sediment. For PFOS and PFOSA there was more of the linear form adsorbed to the solid phases than found in water (PFOS 70-85% vs 40-43%; PFOSA: 93-95% vs 64-75%). For PFOA the difference was not as large in absolute numbers but calculations still indicated a higher log  $K_{OC}$  for n-PFOA.

In an investigation of three Swedish sites drinking water, untreated and treated waste water and sewage sludge, branched and linear forms of PFOS, PFOSA, PFHxS and PFOA were determined (Filipovic and Berger, 2015). The contribution of n-PFOS was on average 62% in the untreated waste water while in the sludge was 86%. Also for PFOSA a tendency for a higher percentage of the linear form in the sludge as compared to influent water was observed (83% vs 68%). For PFOA >91% was linear in the untreated water and in the sludge no branched forms could be detected. Results for PFHxS were not conclusive since all sludge samples had concentrations below limit of detection.

Also for water cleaning the isomeric profile can be important. Östlund (2015) investigated PFAS removal in pilot scale column experiments using anion exchange resin and granular activated carbon. The adsorption efficiency of branched and linear PFOS, PFOSA and PFHxS was assessed. For the anion exchanger no difference due to isomeric form was found, while for active carbon a lower removal of the branched form was seen. At the end of the study (33000 bed volumes) the difference in adsorption was 86% vs 78% for PFOS. The same numbers were 69% vs 57% for PFOSA and 57% vs 40% for PFHxS.

### *Source tracking*

Comparison of the isomeric forms of PFAS in environmental samples with commercial products may yield insights to the manufacturing origins (Benskin et al., 2010). Chen et al. (2015) used an isomer profiling technique to calculate the relative contributions of various industrial origins for PFOA in a Chinese river. The outcome indicated that the major source was electrochemical fluorination (ECF) (55%), followed by linear telomerisation (41%) and isopropyl telomerisation (4%).

Another field of research is to what extent the isomeric composition in e.g. water reflects the transformation of precursors, and the possibility that branched forms are more prone in this regard (Benskin et al, 2010). Leaching tests of sediments showed that the fraction of n-PFOS in the leachate was much higher than in the water in the field (Chen et al., 2015). The difference between the leaching and field results suggested that PFOS precursors could make a contribution to the linear/branched ratio of PFOS in the natural system.

### *Isomer distribution in humans and biota*

For PFAS four major pathways for direct and indirect exposure can be outlined: ingestion of dust, dietary and drinking water intake, and inhalation of air (Gebbink et al., 2015).

Concerning food, n-PFOS comprised 92% of total PFOS in a general Swedish diet and estimates for PFOSA indicated about 98% of the linear form. PFOS and PFOSA isomer ratios in drinking water were 60% n-PFOS and 58% n-PFOSA. Considering total PFOS exposure, taking all pathways into account, the proportion was estimated as 84% linear and 16% branched. The number could be dependent upon the diet, especially at high fish consumption (Gebbink et al., 2015). Enrichment of the linear form of PFOS in animal tissue is frequently seen (often >90% n-PFOS) e.g. for fish and pelagic organisms as well as mammals such as rats and polar bears (Greaves and Letcher, 2013; Miralles-Marco and Harrad, 2015)

However, in humans the isomeric PFOS profile may be more complicated recent studies suggest (Miralles-Marco and Harrad, 2015). In particular human serum samples can sometimes show a br-PFOS content even higher than in the technical ECF products. It has frequently been put forward that these observations can be explained by either an inaccurate estimation of exposure (e.g., unknown precursors, missing or poorly quantified pathways and/or isomer profiles of PFOS and precursors) or an incomplete understanding of the human pharmacokinetic processes (e.g. biotransformation and elimination kinetics) (Gebbink et al., 2015).

Also exposure to PFAS contamination can alter the isomeric profile in humans. In a study of young women in Uppsala (SE) higher total levels of PFHxS as well as an increased fraction of br-PFHxS in serum were observed for those who had been exposed to contaminated drinking water. Still the percentage of br-PFHxS in serum was lower than that found in water (6 vs 20%) (Gyllenhammar et al., 2015).

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